



Original

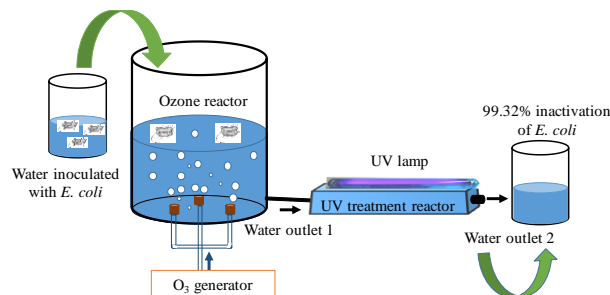
Escherichia coli bacteria inactivation employing ozone and ultraviolet radiation using a reactor with continuously flowing water

A. N. Hernández-Arias^a, B. Jaramillo-Sierra^{a*}, B. G. Rodríguez-Méndez^b, R. Peña-Eguiluz^b,
R. López-Callejas^b, A. Mercado-Cabrera^b, R. Valencia-Alvarado^b, D. Alcántara-Díaz^b

^aTecnológico de Estudios Superiores de Tianguistenco,
Carretera Tenango-La Marquesa km 22, 52650 Santiago Tianguistenco, México

^bInstituto Nacional de Investigaciones Nucleares,
Carretera México-Toluca S/N, 52750 Ocoyoacac, México

Abstract: The study on the inactivation of *Escherichia coli* (*E. coli*) bacteria in water flowing continuously in a treatment system based on two different reactors is presented. The bacteria inactivation is achieved in a water volume of 500 mL inoculated with an *E. coli* concentration of 1.45×10^3 bacteria/mL using combined treatment with ozone and ultraviolet (UV) radiation. Inoculated water firstly is introduced to cylindrical reactor supplied with ozone at a concentration of 1000-1100 ppm and, subsequently, transferred to a rectangular geometry reactor subjected to UV radiation with an energy per volume of 1.44 J/mL. Inactivation efficiency of *E. coli* bacteria in the water of 99.32% was attained in a total treatment time of 180 seconds.



Keywords: Water, *Escherichia coli*, inactivation, ozone, ultraviolet radiation

1. INTRODUCTION

Water is vital for the survival of earth living beings; and for humans, it has multiple uses, among which drinking water stands out. It must have healthy characteristics to be intake; in particular, it must not contain bacteria that can cause diseases. The bacteria have a great adaptation facility

practically to any habitat, and some of them are resistant to high temperatures, extreme dryness, or high humidity (Acquaotta, Ardissino, Fratianni & Perrone, 2017). Several methods of bacteria inactivation have been developed. However, at present, work continues on technologies that can inactivate various microorganisms to offer an adequate quality of water for human consumption (Burns et al., 2012; Murray et al., 2008; Verma, Gupta & Gupta, 2016).

For many years, the classical method of bacteria inactivation has been chlorine addition to water, because

*Corresponding author.

E-mail address: bethsabet.jaramillo@test.edu.mx (B. Jaramillo-Sierra)

Peer Review under the responsibility of Universidad Nacional Autónoma de México.

it is an efficient disinfectant in the inactivation of bacteria, cheap and has a residual effect, that is, it remains active in the water and with the ability to continue inactivating pathogens (Magbanua, Savant & Truax, 2006; Wu, Zhang, You, Yan, & Li, 2016). However, this element in the presence of organic matter can generate toxic compounds such as Trihalomethanes (THMs) and Haloacetic acids (HAAs). Thus, the use of chlorine is restricted in many countries (Hong et al., 2013). As an alternative to chlorine, ozone has been used (formed by three oxygen atoms). The production of this chemical agent is done mainly by electric discharges, through which a gas (oxygen or air) is passed through an electric field, thus generating the ozone in addition of other chemical species. It is an oxidizing agent capable of inactivating bacteria by oxidation of the deoxyribonucleic acid (DNA) (Levén, Wijnbladh, Tuveesson, Kragelund, & Hallin, 2016; Wittmer et al., 2015), lipids of cell membrane, and intracellular organelles (Hashizume et al., 2014). In addition, also to eliminate color, smell, and taste (Loeb, Thompson, Drago, Takahara, & Baig, 2012). The chemical reaction between water and ozone promotes the generation of reactive oxygen species (ROS) such as oxygen (O^+ , O), radicals (O_2^* , $\cdot OH$) and peroxides (H_2O_2) (Burns et al., 2012; Von Gunten, 2003) that cause oxidative stress in bacteria due to its highly oxidizing properties. ROS contain oxygen atoms in their chemical structure that present a wide lifetime range from nanoseconds to hours. ROS are typically generated using photolysis and energy transfer reactions. These processes generate $\cdot OH$ radicals, which are chemical species with redox potential in the range of 1.9 - 2.85 V (Wardman, 2007), which is higher than the ozone redox potential. The presence of $\cdot OH$ has a fundamental role in the elimination of bacteria since it causes physical destruction of the cytoplasmic membrane by oxidation (Burns et al., 2012; Uhm, Choi, Cho & Hwang, 2013).

Likewise, ultraviolet (UV) radiation can inactivate bacteria, viruses, bacterial spores, and parasites. It does neither alter the physical properties of water, nor produce disinfection byproducts because no toxic compounds are added, but it is believed that this method may have a bacteriostatic effect or a repair of the microorganisms allowing again their population (Koivunen & Heinonen-Tanski, 2005). However, this characteristic depends on the type of water to be treated. This method of disinfection acts causing disturbances in the genetic material (DNA)

of the bacteria, which prevents their reproduction. The sufficient wavelength to inactivate the *E. coli* bacterium is in the region UV-C within a range of 200-280 nanometers (nm) with a higher effect at approximately 260 nm (Magbanua, et al., 2006) in liquid. Additionally, other studies inactivate bacteria in the 380-480 nm range (Tomb et al., 2018). Another method of inactivation is based on the use of TiO_2 , which is used as a catalyst and has excellent potential for the disinfection and inactivation of harmful pathogens (Reddy, Kavitha, Reddy, & Kim, 2017). Besides, research is being carried out using the physicochemical and photocatalytic mechanisms of metal oxides (Lebedev, Anariba, Tan, Li, & Wu, 2018). Other procedures that are being carried out in bacterial inactivation, among others the *E. coli*, are employing ozone and photo-assisted disinfection technologies (Gomes, Matos, Gmurek, Quinta-Ferreira, & Martins, 2019).

According to the above, ozone in sequential combination with UV should be an option as a disinfection method, either on surfaces (Kumari et al., 2017) or water (Fang et al., 2014), indicated to be more effective in killing microorganisms than ozone or UV individually. An additional advantage, is the transformation of ozone back into oxygen, without leaving any harmful chemical residue in the water, therefore avoiding the generation of toxic waste (Murray et al., 2008; Karaca & Velioglu, 2014). So that, this technology is expected to become more widespread because of their advantages against other methods as chemical, with public health effects (Magbanua et al., 2006), or catalysis with long exposure periods for achieving reasonable results of inactivation.

In this study, the results of the inactivation of *E. coli* bacteria in water using ozone and UV radiation are presented. After a treatment time of three minutes, it was achieved bacteria inactivation percentages of 99.32% from an initial concentration of 1.45×10^3 bacteria/mL inoculated in a vessel containing 500 mL of water which is associated to a continuous flow disinfection system.

2. EXPERIMENTATION

2.1 EXPERIMENTAL SETUP

Figure 1 shows the experimental diagram for the inactivation of *E. coli* bacteria in water employing ozone and UV radiation. This setup is mainly composed of a cylindrical glass reactor with a volume of ~ 1178 cm³. The lid and the base of this reactor were made of nylamid, and

has two inlets and outlets, one pair of them for water and the other pair for ozone. The last is produced by an ozone generator and injected using three diffusers placed at 120° from each other at the base of the reactor and where the precursor gas is the ambient air. This system has an energy consumption of 10 W. Moreover, a shortwave ultraviolet radiation (UV-C) generated by a 12 W low-pressure mercury lamp made of quartz is radiated to the UV treatment reactor through a quartz cap located at the top, where the distance from the UV lamp to the surface of the water is ~ 2 cm. This last reactor has a rectangular prism shape rounded at the edges built in glass. It has a volume ~ 550 cm³, an inlet and an outlet of water. Meanwhile, the liquid is pumped with an AC water pump from the vessel and by gravity, the water attains the UV treatment reactor and, finally, the liquid is propelled again to the vessel with another pump. This arrangement represents a closed-loop circulating water circuit and when all the water volume is deposited in the vessel, a treatment cycle was done.

The experiment consisted of using 500 mL of sterilized and inoculated water with a concentration of the order of 10³ bacteria/mL, which is deposited in the container, then pumped to the treatment reactor. During water treatment at the first reactor outlet, a sample is acquired to perform the microbiological analysis while a second sample is taken at the outlet of the UV radiation reactor to perform the corresponding microbiological analysis.

Radiation energy per volume (dose) can calculate per unit volume of the liquid to be processed to express the dose (D) in J/L or J/mL as follows (Keyser, Müller, Cilliers, Nel, & Gouws, 2018):

$$D = \frac{P}{q} \quad (1)$$

where P is the maximum power of the UV lamp in watts, and q is the volumetric flow of the liquid in L/s or mL/s.

2.2 CELL CULTURE

To carry the study out the *E. coli* American Type Culture Collection (ATCC) 8739 bacterium was used. It was incubated in 5 mL of Luria-Bertani (LB) nutrient medium for 20 hours at a temperature of 37° C. After incubation, the nutrient medium was withdrawn from the culture tube using a Labnet brand centrifuge operating at 5000 rpm for 10 minutes. After that, the resulting culture was suspended in 5 mL of sterilized water. It was made successive dilutions from this solution to reduce the concentration, until being able to count the number of bacteria employing a Carl Zeiss brand phase contrast microscope and a Neubauer chamber. Then a concentration of the order of 10³ bacteria/mL was inoculated in 500 mL of water and deposited in a vessel to start the treatment.

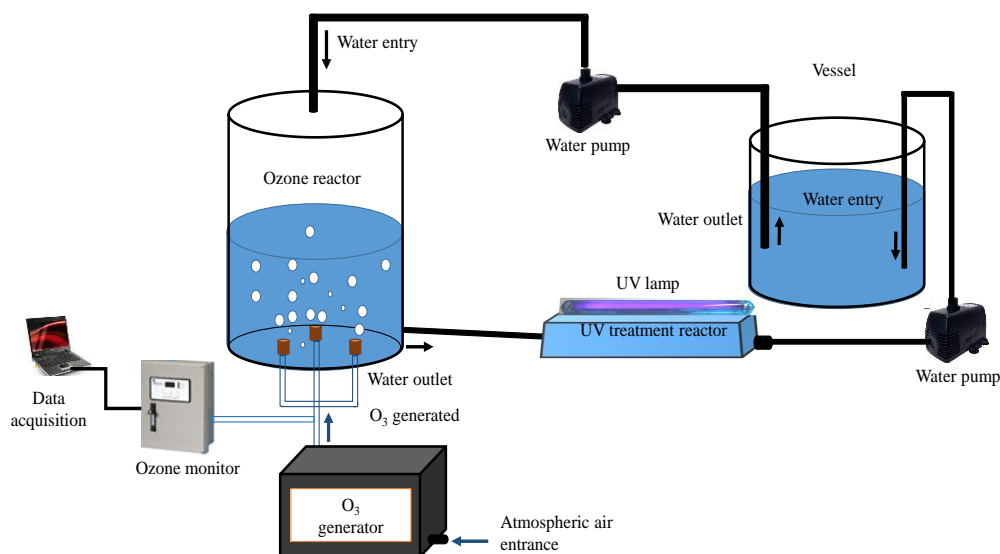


Fig. 1. Experimental setup for the inactivation of *E. coli* bacteria in water.

2.3 MICROBIOLOGICAL ANALYSIS

The water inoculated with bacteria *E. coli* and subsequently treated by ozone and UV radiation phases were subjected to a microbiological analysis to determine the results of bacteria deactivation. Afterward, three 0.1 mL aliquots obtained from treated samples were placed on previously prepared Petri dishes with LB solid nutrient medium, later dispersed and incubated at a temperature of 37° C for 20 hours.

The reference sample was taken from the vessel at the beginning of the water treatment process. Considering that each viable bacterium forms a colony (CFU), the colonies were counted after incubation using a SOLBAT colonies counter. This process is applied to the treated samples, so after counting each one of the CFUs, the number of surviving bacteria (*NBS*) in CFU/mL is obtained from the following expression:

$$NBS = \frac{NC}{VMU} \quad (2)$$

where *NC* is the number of counted colonies (in CFU) and *VMU* is the volume of the used sample (0.1 mL).

While the inactivation percentage is calculated as:

$$\% \text{ Inactivation} = 100\% - \frac{NBS}{No} \times 100\% \quad (3)$$

where: *No* is the number of counted colonies (CFU/mL) of the reference sample.

3. RESULTS

3.1 OZONE CHARACTERIZATION

The characterization of the ozone was made at the entrance of the treatment reactor using an ozone monitor (Teledyne Instruments 460L) through an interface to a personal computer using a data acquisition system. Results were obtained by sampling every second and plotted later. The amount of ozone supplied to the reactor at steady-state (approximately 60 s after ozone generator is turned-on) was in the range of 1000-1100 ppm. Figure 2 shows the resulting ozone concentration at the treatment reactor inlet.

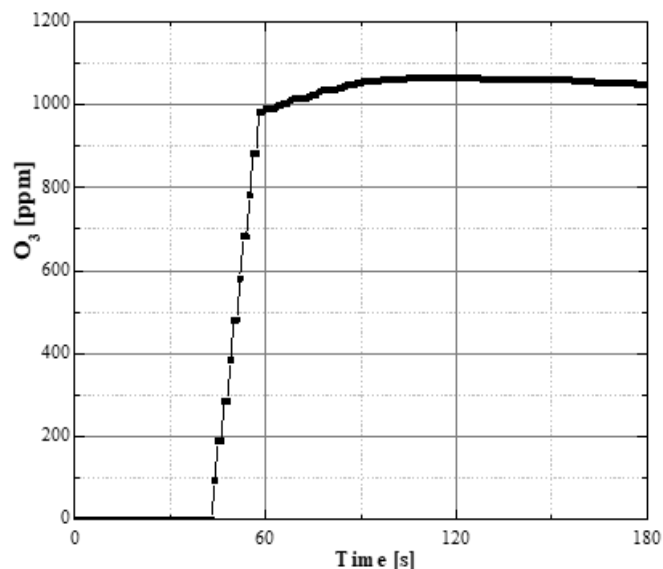


Fig. 2. Ozone concentration at the entrance of the treatment reactor.

3.2 CHARACTERIZATION OF UV RADIATION

It can be observed in Figure 3 obtained optical emission spectrum from the UV lamp used. It was acquired in a wavelength range of 200-350 nm, showing characteristic peaks over wavelengths of 265-335 nm with peak emission at 254 nm, which corresponds to the typical wavelength where bacteria show susceptibility to inactivation (Coohill & Sagripanti, 2008). The characterization was achieved using a Jaz OceanOptics™ optical emission spectroscopy (OES) with an optical resolution of 0.3 nm. The power of the lamp was 12 W, and the volumetric flow was 8.33 mL/s, then, using (1) the estimated UV radiation energy per volume was 1.44 J/mL.

3.3 INACTIVATION OF *E. COLI* BACTERIA

Figure 4 shows the qualitative results of *E. coli* bacteria on agar after 20 hours of incubation. Figure 4.a corresponds to the reference sample representing a concentration of 1.45×10^3 CFU/mL. Figure 4.b and Figure 4.c show the results after the first treatment cycle for treatments of ozonation and ozonation and UV radiation, respectively. Figure 4.d and Figure 4.e depict the results after the second treatment cycle for treatments of ozonation and ozonation and UV radiation respectively, and, Figure 4.f and Figure 4.g correspond to the results after the third treatment cycle for both applied treatments.

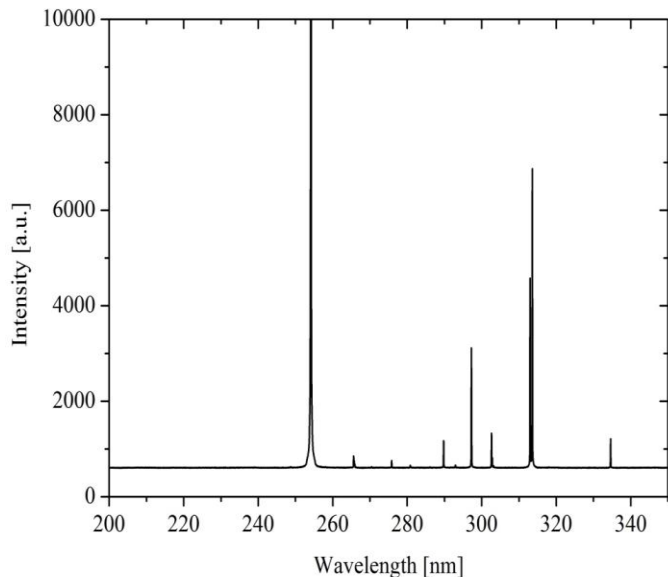


Fig. 3. Characterization of the ultraviolet radiation lamp.

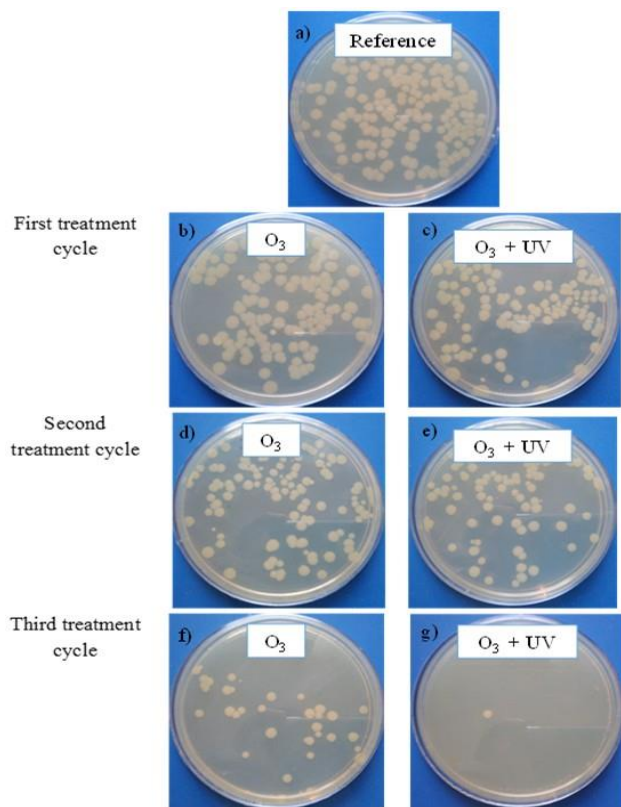


Fig. 4. Inactivation of *E. coli* using ozone and UV radiation. Images of experimental results after 20 h incubation.

The inactivation of the *E. coli* bacteria was evaluated from the comparison of CFU/mL of the reference sample Figure 4(a) concerning the CFU/mL after performing the combined ozonation and UV radiation treatments. From

the results shown in Figure 5, it is observed that the average initial concentration about 1.45×10^3 CFU/mL, after the first treatment cycle bacteria concentration was reduced to 920 CFU/mL with ozone treatment and to 860 CFU/mL with ozone treatment plus UV radiation. Afterward, the water was recirculated, and after the second treatment cycle, bacteria concentration was reduced to 760 CFU/mL with ozone and to 620 CFU/mL with ozonation coupled to UV radiation. Finally, the resulting bacteria concentration at the end of the third cycle was 250 CFU/mL with ozone and 10 CFU/mL applying ozone and UV radiation.

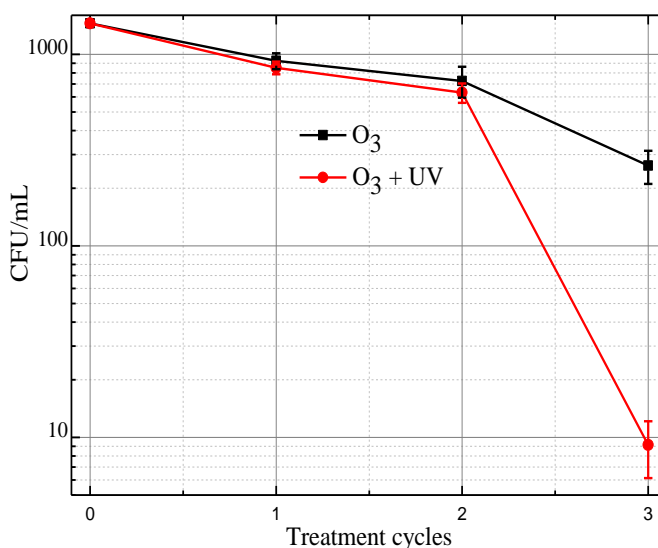


Fig. 5. Inactivation of bacteria in function of treatment cycles.

The percentages of *E. coli* bacteria inactivation are depicted in figure 6, where it can be observed that as increasing the number of treatment cycles, the higher percentage of bacteria inactivation was attained. Therefore, at the end of the first treatment cycle, the obtained bacteria inactivation with O₃ and combined ozone with UV radiation were respectively 37% and 41%. While after the second treatment cycle, the resulting deactivation efficiencies were 48% with O₃ and 58% with O₃ and UV radiation. Finally, the higher efficiency percentages were attained the third treatment cycle (180 s), afterward with 83% supplying O₃ and 99.32% combining ozone and UV radiation. These results indicate that although both processes can be used in the inactivation of *E. coli*, ozone with UV has the potential to enhance the disinfection process.

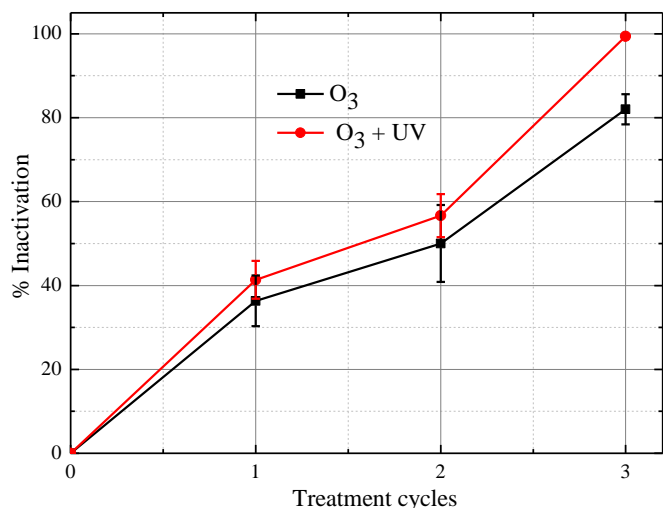


Fig. 6. *E. coli* bacteria inactivation percentage as a result of the application of ozone and combined ozone and UV radiation.

4. DISCUSSION

When ozone is applied to a liquid, a large number of bubbles are generated as has been observed by other authors (Sumikura, Hidaka, Murakami, Nobutomo, & Murakami, 2007; Mishra et al., 2017). In our case, the ozone supply into the ozonation reactor promotes a considerable amount of bubbles and microbubbles at the whole volume of contained water, achieving a high and uniform contact between generated ROS with bacteria inoculated in water. ROS can be produced in this process by the next chemical reactions. Ozone is broken down into atomic and diatomic oxygen by the reaction:



Atomic oxygen can take as a vehicle the generated bubbles and has a more effective interaction with the bacteria cellular membrane promoting their lipid oxidation, abstracting hydrogen from membrane lipids and proteins to form peroxides (Wang, Libardo, Angeles-Boza, & Pellois, 2017). Also, when atomic oxygen reacts with water, it generates $\cdot\text{OH}$ radical (Karaca & Velioglu, 2014) which has a considerable redox potential in the range of 1.9 - 2.85 V, which can interact directly with bacteria. ROS such as the $\cdot\text{OH}$, as observed in previous studies (Rodríguez-Mendez et al., 2013), could also be the cause of a chemical attack on the bacteria, such as ozone that generates oxidative

stress on the cellular membrane (Karaca & Velioglu, 2014; Uhm, et al., 2013). The $\cdot\text{OH}$ is generated by:



In the first instance, this mechanism could be responsible for the inactivation effectiveness against *E. coli* by means ozone. Furthermore, ozonation treatment is complemented with UV radiation, because it generates $\cdot\text{OH}$ radicals in the presence of water at $\lambda = 200\text{-}280$ nm. As it is established in the next reaction:



In addition to chemical oxidation processes with the membrane cells produced by the generated of chemical species with a more significant redox potential such as: O, O₃, $\cdot\text{OH}$; ones UV radiation can penetrate the bacteria making a muddle of its genetic material which causes the inactivation of *E. coli* bacteria (The above mentioned could explain the results shown in Fig. 5. and Fig. 6, where the effect of the O₃/UV treatment (similar to O₃ until the second treatment cycle) against *E. coli* from the third treatment is evident. As can be seen, in overall water treatment of three cycles (180 seconds), *E. coli* were inactivated much more rapidly with the combined ozonation and UV treatments as compared to the O₃ treatment, corresponding to more than an order of magnitude of inactivation higher than with O₃ alone. These characteristics made this a low cost and effective method to inactivate bacteria in water flowing continuously with high bacterial inactivation efficiency, in relatively short treatment times.

5. CONCLUSIONS

The here proposed treatment system turned out to be very efficient, obtaining a 99.32% inactivation of *E. coli* bacteria at a concentration of 1.45×10^3 CFU/mL. Two inactivation techniques were applied in the treatment system: the first one based on the injection of ozone at the base of a cylindrical reactor and the second one determined by UV radiation exposure in a rectangular reactor, achieving a satisfactory efficiency magnitude. It must be noted that results mainly are determined by the biological interaction between bacteria structure and the

generated chemical species and applied UV radiation. This methodology does not make use of supplement chemical products and consumes ambient air as a primary resource, making this proposal harmless to the environment.

ACKNOWLEDGMENT

This research was supported by COMECYT, Mexico. The authors thank A. Almazan V., G. Colin V. and E. J. Salgado M. for their technical support.

REFERENCE

- Acquaotta F., Ardissino G., Fratianni S. & Perrone M. (2017). Role of climate in the spread of shiga toxin-producing *Escherichia coli* infection among children. *International Journal of Biometeorology* 61(9), 1647-1655.
- Burns, J. M., Cooper, W. J., Ferry, J. L., King, D. W., DiMento, B. P., McNeill, K., ... & Rose, A. L. (2012). Methods for reactive oxygen species (ROS) detection in aqueous environments. *Aquatic Sciences*, 74(4), 683-734.
- Coohill, T. P., & Sagripanti, J. L. (2008). Overview of the inactivation by 254 nm ultraviolet radiation of bacteria with particular relevance to biodefense. *Photochemistry and photobiology*, 84(5), 1084-1090.
- Fang, J., Liu, H., Shang, C., Zeng, M., Ni, M., & Liu, W. (2014). *E. coli* and bacteriophage MS2 disinfection by UV, ozone and the combined UV and ozone processes. *Frontiers of Environmental Science & Engineering*, 8(4), 547-552.
- Gomes, J., Matos, A., Gmurek, M., Quinta-Ferreira, R., & Martins, R. (2019). Ozone and Photocatalytic Processes For Pathogens Removal from Water: A Review. *Catalysts*, 9(1), 46.
- Hashizume, H., Ohta, T., Takeda, K., Ishikawa, K., Hori, M., & Ito M. (2014). Oxidation mechanism of *Penicillium digitatum* spores through neutral oxygen radicals. *Japanese Journal of Applied Physics*, 53(1), 010209.
- Hong, H., Xiong, Y., Ruan, M., Liao, F., Lin, H., & Liang, Y. (2013). Factors affecting THMs, HAAs and HNMs formation of Jin Lan Reservoir water exposed to chlorine and monochloramine. *Science of the Total Environment*, 444, 196-204.
- Karaca, H., & Velioglu, Y. S. (2014). Effects of ozone treatments on microbial quality and some chemical properties of lettuce, spinach, and parsley. *Postharvest Biology and Technology*, 88, 46-53.
- Keyser, M., Müller, I. A., Cilliers, F. P., Nel, W., & Gouws, P. A. (2008). Ultraviolet radiation as a non-thermal treatment for the inactivation of microorganisms in fruit juice. *Innovative Food Science & Emerging Technologies*, 9(3), 348-354.
- Koivunen, J., & Heinonen-Tanski, H. (2005). Inactivation of enteric microorganisms with chemical disinfectants, UV irradiation and combined chemical/UV treatments. *Water research*, 39(8), 1519-1526.
- Kumari D., Kumar R. N., Bhaskar V., Chakravarthy D. P. & Sudershan V. R. (2017). Efficacy of Ozone in Combination with UV for Inactivation of Selective Foodborne Pathogens. *Acta Scientifica Nutritional Health* 1, 38-42.
- Reddy, P. V. L., Kavitha, B., Reddy, P. A. K., & Kim, K. H. (2017). TiO₂-based photocatalytic disinfection of microbes in aqueous media: a review. *Environmental research*, 154, 296-303.
- Lebedev, A., Anariba, F., Tan, J. C., Li, X., & Wu, P. (2018). A review of physiochemical and photocatalytic properties of metal oxides against *Escherichia coli*. *Journal of Photochemistry and Photobiology A: Chemistry*, 360, 306-315.
- Levén, L., Wijnbladh, E., Tuveesson, M., Kragelund, C., & Hallin, S. (2016). Control of *Microthrix parvicella* and sludge bulking by ozone in a full-scale WWTP. *Water Science and Technology*, 73(4), 866-872.
- Loeb, B. L., Thompson, C. M., Drago, J., Takahara, H., & Baig, S. (2012). Worldwide ozone capacity for treatment of drinking water and wastewater: a review. *Ozone: Science & Engineering*, 34(1), 64-77.
- Magbanua, B. S., Savant, G., & Truax, D. D. (2006). Combined ozone and ultraviolet inactivation of *Escherichia coli*. *Journal of Environmental Science and Health, Part A*, 41(6), 1043-1055.
- Mishra, N. S., Reddy, R., Kuila, A., Rani, A., Mukherjee, P., Nawaz, A., & Pichiah, S. (2017). A review on advanced oxidation processes for effective water treatment. *Current World Environment*, 12(3), 470.
- Murray, B. K., Ohmine, S., Tomer, D. P., Jensen, K. J., Johnson, F. B., Kirsi, J. J., ... & O'Neill, K. L. (2008). Virion disruption by ozone-mediated reactive oxygen species. *Journal of virological methods*, 153(1), 74-77.
- Rodriguez-Mendez, B. G., Hernandez-Arias, A. N., Lopez-Callejas, R., Valencia-Alvarado, R., Mercado-Cabrera, A., Pena-Eguiluz, R., ... & De La Piedad-Beneitez, A. (2013). Gas flow effect on *E. coli* and *B. subtilis* bacteria inactivation in water using a pulsed dielectric barrier discharge. *IEEE Transactions on Plasma Science*, 41(1), 147-154.
- Sumikura, M., Hidaka, M., Murakami, H., Nobutomo, Y., & Murakami, T. (2007). Ozone micro-bubble disinfection method for wastewater reuse system. *Water Science and Technology*, 56(5), 53-61.

- Tomb, R. M., White, T. A., Coia, J. E., Anderson, J. G., MacGregor, S. J., & Maclean, M. (2018). Review of the Comparative Susceptibility of Microbial Species to Photoinactivation Using 380–480 nm Violet-Blue Light. *Photochemistry and photobiology*, *94*(3), 445-458.
- Uhm, H. S., Choi, E. H., Cho, G. S., & Hwang, D. H. (2013). Influence of reactive oxygen species on the sterilization of microbes. *Current Applied Physics*, *13*, S30-S35.
- Verma, K., Gupta, D., & Gupta, A. B. (2016). Optimization of ozone disinfection and its effect on trihalomethanes. *Journal of environmental chemical engineering*, *4*(3), 3021-3032.
- Von Gunten U. (2003). Ozonation of drinking water: Part I. Oxidation kinetics and product formation. *Water Research* *37*, 1443-1467.
- Wang, T. Y., Libardo, M. D. J., Angeles-Boza, A. M., & Pellois, J. P. (2017). Membrane oxidation in cell delivery and cell killing applications. *ACS chemical biology*, *12*(5), 1170-1182.
- Wardman, P. (2007). Fluorescent and luminescent probes for measurement of oxidative and nitrosative species in cells and tissues: progress, pitfalls, and prospects. *Free radical biology and medicine*, *43*(7), 995-1022.
- Wittmer, A., Heisele, A., McArdell, C. S., Böhrer, M., Longree, P., & Siegrist, H. (2015). Decreased UV absorbance as an indicator of micropollutant removal efficiency in wastewater treated with ozone. *Water Science and Technology*, *71*(7), 980-985.
- Wu, D., Lu, G., Zhang, R., You, H., Yan, Z., & Li, Y. (2016). Disinfection characteristics of the combined ultraviolet radiation and ozone process using *Escherichia coli* as a probe. *Water Science and Technology: Water Supply*, *16*(1), 163-170.